Guidelines for assessing the valorization of a waste into cementitious material: dredged sediment for production of self compacting concrete

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ABSTRACT: This article presents some guidelines in order to analyse the feasibility of including a waste material in the production of a structural cementitious material. First of all, the compatibility of the waste with a cementitious material has to be assured; then, if necessary, a decontamination step will be carried out; after, decision on the type of material has to be taken based on different aspects, with special emphasis on the granularity. As a last step, mechanical, environmental and durability properties have to be evaluated. Then the procedure is illustrated with a full example, obtaining a self compacting concrete (SCC) including dredged sediment taken from a Spanish harbour.

KEYWORDS: Protocol; Waste treatment; Cementitious material; Dredged sediment; Self-compacting concrete

1. INTRODUCTION

In the last years, lot of research has been done in including wastes into concrete, mainly for secondary uses but also trying to accomplish with structural performance. As an example, a bibliographic search have been done in the Web of Science, looking for “concrete” and “waste” within the topic, in all domains of science and including all the years in the repository until October 2014. Within these results, filters were applied to include different kind of wastes, as are “dredged sediment”, and other wastes more typically used in concrete fabrication, as “industrial”, “agricultural” or “demolition” among others (1–10). The results are given in Figure 1, where it can be deduced that even though there are a lot of contributions in this area, and society concerns about the reutilization of waste in concrete fabrication, the research is mainly focused on the reuse of several wastes, as
industrial, demolition or glass, for most abundant. Only 12 papers were found for the topic including dredged sediments as a waste to be used in concrete fabrication, and no one is found for the use of dredged sediments in self compacting concrete. It is remarkable that no general rules, as established protocols, are found in the bibliography. In this way, there are no previous efforts in order to systematically evaluate and decide on the feasibility for including a waste into a cementitious matrix.

Even though 33 papers were found including the words “waste + concrete + protocol”, after its examination, only one tried to reflect a real protocol (11), but it was based in the reuse of demolition materials, nothing to do with the scope of this paper.

It is also remarkable that the research in the area of reusing materials for self compacting concrete (SCC) started in the last decade (12–14), and has increased very fast in the last years with interesting contributions (1–3, 15–24), so, the necessity of some guidelines to assess the inclusion of these wastes in SCC is clear.

Filling this gap is the objective undertaken in this paper, which presents general guidelines to be followed in order to analyse the feasibility of including a waste material in the production of one of the most widely used building materials: concrete.

These guidelines have been illustrated on the basis of using dredged material in performing SCC. This is an interesting case study, as no bibliography has been found on it. Dredged sediment has been chosen because dredging is necessary to create and maintain navigation channels in naval facilities, and despite several hundred million cubic meters of sediment dredged each year just in U.S. most of this dredged material is disposed in open water, confined disposal facilities, and upland disposal facilities, not having so much research dedicated to, as reflected in the bibliographic results showed in Figure 1. In addition, in U.S., consistent with the National Dredging Policy, EPA encourages the “beneficial reuse” of dredged material. In Europe, having the same problem, according to new European Union Directives, mainly the EU Water Framework Directive, it is necessary to find an environmental solution to these materials. Dredged sediments have to be seen as a valuable resource better than a waste.

The reason for choosing SCC as the product to be constructed has been derived from the characteristics of the waste, taking into account its special characteristics in relation with their intrinsic environmental friendly technology: is able to flow by its own weight, eliminating the need of vibration, which can be translated into substantial reduction in energy, labour cost and construction time. Additionally, it contributes to a better working environment by eliminating the impact of noise and vibration as well as a making more comfortable the periods of works for the affected society.

2. GUIDELINES FOR ASSESSING THE VALORISATION OF THE WASTE

The scheme proposed and designed to assess the feasibility of a waste material to be incorporated in a cementitious material is given in Figure 2.

- Step 1: First of all, the waste material has to be subjected to a thorough chemical and mineralogical analysis in order to establish their elemental composition compatibility with a cementitious material. This is a very important step in the process, as if compatibility is not clear, it will be discarded for this type of valorisation.
- Step 2: If the composition of the waste is compatible with cement, hazardous characterization
of the waste has to be carried out. According to the definition, characteristic hazardous or toxic wastes are determined by evaluating their ignitability, corrosivity, reactivity, and toxicity. If they do not accomplish with these requirements, the following question is: it is possible to remediate it? This is quite important in the case of toxicity, with wastes having heavy metals and organic pollutants, which is the most frequent case. Different techniques to remediate contaminated wastes can be found in literature, for example, electrokinetic remediation (25–27). The same apply for components non hazardous but deleterious for the use: for example chlorides for reinforced concrete (28–30).

- **Step 3:** As a third step, it is necessary to carry out a thoroughly granulometric characterization.
- **Step 4:** Having material and hazardous compatibility assured, in function of composition and granulometry characteristics, the most suitable material to be produced, that includes it, has to be established to be decided.
- **Step 5:** The following step is the design of the material chosen with traditional components and including the waste, trying always to introduce the highest amount of waste compatible with the expected use.
- **Step 6:** After having designed the initial mix, according to the results obtained in the previous steps, the specimens have to be analysed both in fresh and in hardened state, where mechanical and environmental characteristics of the concrete, together with its durability properties have to be evaluated.

If results of step 6 are satisfactory, the material designed is feasible. If not, it is necessary to go again to step 4 and reconsider the type of material to fabricate.

**3. EXPERIMENTAL PROCEDURE: APPLICATION TO THE DREDGED MATERIAL**

**3.1. Characterization of the dredged materials**

The dredged material used in this research has been real non polluted marine dredged sediment taken from a Spanish Mediterranean harbour. After dredging, the material was stored under controlled conditions. Previous works about the characterization of these materials had been taken into account (31–34).

The composition of the dredged sediment used in this investigation was characterized on elemental composition by chemical analysis. The mineralogical determination was carried out by X-Ray diffraction.

Metal determination was carried out by inductively coupled plasma atomic emission spectroscopy (ICP-AES), also referred to as inductively coupled plasma optical emission spectrometry (ICP-OES), has been use on this propose. ICP-AES is an analytical technique used for the detection of trace metals in the environment.

ICP-AES allows detecting metals in all matrix, including soils water, aquosus samples, sediments, but a previous digestion is required (Al, Ba, B, Cu, Fe, Mn, Zn, Cd, Ni, Pb, Ca, Mg, K, Na, etc.).
For the measurements of metals like As, Hg, Sn, Se, Sb and Bi, a coupled hydride generation is used.

All the analysis were carried out by an external accredited laboratory following a regular protocol. The detection limits of the techniques are shown in Table 1.

For the ICP-AES determinations, a high frequency generator feeds an induction coil that produces a strong magnetic field. Argon flow is introduced into this magnetic field through a quartz torch located inside the coil. Argon is heated and ionized to form plasma (ionized gas sufficiently to conduct electrical energy).

A pneumatic nebulizer transforms liquid sample into an aerosol that is injected to the centre of the inductively coupled plasma (ICP). At the temperature reached (10,000 K) many atoms are excited and emit a characteristic spectrum at wavelengths determined for each element wave.

All the analysis for sediment physical and chemical characterization were made according to Spanish recommendations for dredged materials and following the recommended protocols (31). The dry weight fraction was determined by weight loss at 105 °C. For the rest of analyses, sediments were dried at 40 °C for 24 h.

Additionally, physical characterization (granulometric analyses) and determination of the total organic content (TOC) was carried out. Grain size distribution followed UNE 103 101 and total organic carbon (TOC) content was estimated by loss of ignition (LOI, determined heating the sample during 15 minutes at the temperature of 950 °C±25 °C) and gravimetric determination as recommended for small dredged volumes and applying the following expression to express the results as total organic carbon.

\[
\text{TOC (g kg}^{-1}) = 0.35 \text{LOI (g kg}^{-1})
\]

It was analysed in order to assess the presence of heavy metals, polycyclic aromatic hydrocarbons (PAHs), organophosphorous pesticides and other organic chlorinated compounds. These analyses are made using the fraction with particle size lower than 63 μm.

### 3.2. Fabrication of the concrete specimens

Concrete was cast with cement type I 42.5R without additions. Other components, apart from the sediment were water, sand and coarse aggregates. Three different mixes were tested. Its composition is shown in Table 2.

### 3.3. Characterization of the concrete

In fresh state, the fluency of the concrete was characterized through the slump test using the Abram’s cone and the consistency of the concrete was measured following the UNE EN 12350-2 standard (35) by measuring the spread of the concrete on a flat plate subjected to a shocking movement. It was also followed the UNE EN 12350-8 standard (36), suitable for SCC. The density of the fresh concrete was measured according to the UNE EN 12350-6 standard (37). Another important parameter is the entrapped air of the concrete, which was determined through by means of the standard UNE EN 12350-7 (38).

In hardened state it was analyzed from the point of view of its mechanical, environmental and durable properties: From the mechanical point of view, compressive strength tests were carried out following the standard UNE-EN 12390-3 (39).

From an environmental point of view, leaching tests were carried out following the standard EN 12457-2:2002 (40). This test involves immersing the sample at a specified leachant volume in order to extract the components of the material. The solid residue is separated by filtration and an aliquot of the leachate is taken and analysed following the standards UNE ENV 12506 (41) and UNE ENV 13370 (42).

The durability properties were evaluated by analysis of porosity, pore size distribution and density

| Table 1. Detection limits for the techniques used for metal composition (ICP-AES) and mineralogical composition (chemical analysis) |
|---|---|---|
| Metal composition (Detection limits, mg/kg) | Mineralogical composition (Detection limits, %) |
| As | 0.1 | SiO₂ | 0.02 |
| Hg | 0.1 | Al₂O₃ | 0.1 |
| Cd | 2 | Fe₂O₃ | 0.1 |
| Cr | 2 | CaO | 0.1 |
| Co | 2 | MgO | 0.1 |
| Ni | 2 | Na₂O | 0.01 |
| Pb | 10 | K₂O | 0.01 |
| Zn | 2 | SO₃ (Sc) | 0.01 |

| Table 2. Details of the mixes composition |
|---|---|---|
| Mix | A | B | C |
| Cement I 42.5 R | 420 | 450 | 450 |
| Dredged material | 73 | 80 | 100 |
| water | 205 | 202 | 202 |
| w/c | 0.49 | 0.45 | 0.45 |
| Silicious coarse aggregate (6/12) | 717 | 676 | 667 |
| Silicious sand (0/6) | 850 | 870 | 859 |
| Admixture I | 9.24 | 7.1 | |
| Admixture II | 7.425 | | |
in two specimens by mercury intrusion porosimetry (MIP), using a Micromeritics 9320 Porosimeter. Also, the crystalline phases of the concrete have been analysed in duplicate powdered samples by X-Ray diffraction using a diffractometer model D8 Advance of Bruker AXS. Other indicator of durability is the electrical resistivity, which was measured by means of a commercial resistivimeter. Additionally, two different experiments to evaluate transport properties through the matrix were carried out: the first one was the measurement of the steady and non-steady state chloride diffusion coefficients and the other was capillary absorption. The steady (Ds) and non-steady (Dns) state chloride diffusion coefficients were determined using the multi-regime method (30) standardized as UNE 83987-2009 (43). The water take-up of hardened concrete was determined by capillary absorption. This method was standardized as UNE 83982:2008 (44).

All the procedures followed can be found in the mentioned bibliographic references.

4. RESULTS AND DISCUSSION

Step 1: Chemical and mineralogical characterization of the dredged material

Elemental composition of the original sediment, and on the fractions smaller than 2 mm and smaller than 63 μm (sand + fine and fine fractions) are given in Table 3, where it can be seen that the major component is silicon oxide followed by calcium carbonate in all the different fractions analysed. Loss of material by calcination (LOI, lost of ignition) was also determined heating the sample during 15 minutes at the temperature of 950 °C±25 °C.

This composition has been confirmed by X-ray diffraction, where presence of quartz and calcite are the main components. Silicates of Fe and Mg are also present, dolomite, Fe and Ti sulphate, complex chlorides including different species, iron, and some clay (chamosite) in small proportions have also been identified. Expansive clays of the group of phyllosilicates (montmorillonite, saponite, nontronite, bentonite sepiolite and palygorskite), have been looked for specifically, not having been identified in the sediment.

Therefore, the sediment is compatible with cementitious materials.

Concerning pH of the sediment, 1 g of sediment suspended in 100 ml deionized water, gives a pH in the solution of 8.5, which, as expected, implies alkaline pH compatible with cementitious materials.

Therefore, it is possible to go ahead to step 2.

Step 2: Hazardous characterization of the waste

The dredged sediment is a mineral waste that is not ignitable, corrosive or reactive. So, the point to analyse from a hazardous point of view is the toxicity. There are different recommendations and environmental limits concerning discharge of dredged sediments to the sea. According to the Spanish recommendations for the management of dredged material in the Spanish harbours (31), toxicologically, the dredged sediments can be divided into three categories in function of their relative values concerning two levels of action: level 1 and level 2.

Category I: their effects on the marine flora and fauna are null or negligible. Therefore, these sediments can be discharged into the sea considering only mechanical effects. For this category, the level of contaminants is smaller than the level of action 1.

Category II: Moderated amount of contaminants. They could be discharged into the sea in a controled way. For this category, the level of contaminants is higher than the level of action 1 but smaller than the level of action 2.

Category III: Dredged materials highly contami-nated. They must be isolated from the sea or reme-diated accordingly. For this category, the level of contaminants is higher than the level of action 2.

In terms of toxic substances contained in the material, the chemical characterization of pollutants was done over the fraction smaller than 63 μm, where they accumulate. The results obtained by ICP-AES, as well as the level of action 1 and 2 specified in (31) are given in Table 4, were it can be seen that the sediment corresponds to Category I.

Table 3. Characterization of the main components of the sediment. Loss of material by calcination/ignition (LOI −15 minutes at the temperature of 950 °C±25 °C)

<table>
<thead>
<tr>
<th>Component</th>
<th>Original sediment</th>
<th>&lt;2 mm</th>
<th>&lt;63 μm</th>
</tr>
</thead>
<tbody>
<tr>
<td>SiO₂</td>
<td>70.6</td>
<td>71.9</td>
<td>68.5</td>
</tr>
<tr>
<td>Al₂O₃</td>
<td>5.1</td>
<td>4.7</td>
<td>5</td>
</tr>
<tr>
<td>FeO</td>
<td>2.3</td>
<td>2.2</td>
<td>2.0</td>
</tr>
<tr>
<td>CaO</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>CaCO₃ %</td>
<td>26.3</td>
<td>27.4</td>
<td>34.2</td>
</tr>
<tr>
<td>MgO</td>
<td>2.3</td>
<td>2.6</td>
<td>2.1</td>
</tr>
<tr>
<td>Na₂O</td>
<td>1.5</td>
<td>1.2</td>
<td>0.93</td>
</tr>
<tr>
<td>K₂O</td>
<td>2.0</td>
<td>1.1</td>
<td>1.6</td>
</tr>
<tr>
<td>TiO₂</td>
<td>0.24</td>
<td>0.24</td>
<td>0.43</td>
</tr>
<tr>
<td>SO₂</td>
<td>0.11</td>
<td>0.07</td>
<td>0.11</td>
</tr>
<tr>
<td>Loss of material by calcination</td>
<td>16</td>
<td>16</td>
<td>20</td>
</tr>
<tr>
<td>Humidity</td>
<td>21.9</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Solid concentration</td>
<td>78.1</td>
<td>4.1</td>
<td></td>
</tr>
</tbody>
</table>
The granulometric distribution of the dredged sediment has been carried out according to UNE 103-102, by sedimentation, and is illustrated in Figure 3. The deposits belong preferably to the sand fraction between 2 mm and 63 μm (94%). The clay and silt fraction reaches 5%. In this figure, the values obtained in the granulometric analysis of the sediment are also compared with the limit values of the fine aggregates for concrete (45) and that of the filler for SCC concretes. From this figure, it can be deduced that the sediment is finer than the fine arid limits, not being finer enough as to be filler for SCC.

**Step 4: Decision on the most suitable material, including the waste, to be produced**

Considering the composition of the sediment, but mainly their granulometric characteristics (beyond the limits of fine aggregates) it has been chosen as a first option, trying to make self compacting concrete, SCC. This decision has also been motivated for the novelty, as seen in the bibliographic search, and for the high added value that it would suppose.

**Step 5: Design of the material**

First of all, it has to be pointed out that the intention of this step is not to optimise the mix of SCC, but to demonstrate the feasibility of their production with this waste on the basis of their properties. Therefore, provided the granulometry of the sediment, it could be used as a partial substitution of fine aggregates or as quasi-filler, with addition of more fines.

**4.1. Trials using the sediment as substitution of fine aggregate**

The design of the sample started replacing 8% of the fine aggregate with dredged sediment, in relation with a conventional concrete (mix A in Table 2). The cement type used was I 42.5 R without additions. The particle size was 6/12 for the coarse aggregate and 0/6 for the sand. After several trials for optimization, the concrete obtained contained 420 Kg/m³ of cement with a water/cement ratio of 0.49 and a consistency of 12. In fresh state the behaviour of this concrete was good, however, owing to the high content of fine arid and therefore its high demand of water, it was not so cohesive and not very fluid (see Figure 4). This behaviour was confirmed with the data obtained in the compressive strength measured at the age of 28 days. These data show an average value of 34 MPa. Without using the sediment this value is close to 40 MPa.

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**Table 4. Toxicological characterization of the dredged material. Results are given by mg/Kg of dry material**

<table>
<thead>
<tr>
<th>mg/Kg</th>
<th>&lt;63 μm</th>
<th>Level 1</th>
<th>Level 2</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mercury</td>
<td>0.21</td>
<td>0.6</td>
<td>3.0</td>
</tr>
<tr>
<td>Cadmium</td>
<td>1.5</td>
<td>1.0</td>
<td>5.0</td>
</tr>
<tr>
<td>Lead</td>
<td>11.0</td>
<td>120</td>
<td>600</td>
</tr>
<tr>
<td>Copper</td>
<td>8.5</td>
<td>100</td>
<td>400</td>
</tr>
<tr>
<td>Zinc</td>
<td>150</td>
<td>500</td>
<td>3000</td>
</tr>
<tr>
<td>Chromium</td>
<td>30.5</td>
<td>200</td>
<td>1000</td>
</tr>
<tr>
<td>Nickel</td>
<td>9.5</td>
<td>100</td>
<td>400</td>
</tr>
<tr>
<td>PCB (28, 52, 101 118, 138, 153 y 180 IUPAC)</td>
<td>&lt;0.03</td>
<td>0.03</td>
<td>0.1</td>
</tr>
<tr>
<td>Arsenic</td>
<td>6.6</td>
<td>80</td>
<td>200</td>
</tr>
<tr>
<td>Total PCB</td>
<td>&lt;0.03</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Oils and fats</td>
<td>502</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Chlorinated organic pesticides</td>
<td>&lt;0.05 (each)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Organic extractable compounds</td>
<td>&lt;0.05 (each)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Total polycyclic aromatic hydrocarbons (PAHs)</td>
<td>&lt;0.05 (each)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Organophosphorus pesticides</td>
<td>&lt;0.5 (each)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Tin organic compounds</td>
<td>&lt;0.01 (each)</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Total petroleum hydrocarbons</td>
<td>&lt;5</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>
4.2. Trials using the sediment as filler

According to the results obtained in the preliminary tests, the possibility of reusing the sediment as a part of the filler instead of using it in the replacement of fines was evaluated.

Several tests were done with the same siliceous aggregate previously used, 450 Kg of cement with a water/cement ratio of 0.45 and 80 Kg of dredged material (mix B in Table 2). The behaviour observed for this mixture was very close to the fluidity, with the consistency of a self compacting concrete but without achieving the typical characteristics of this material since the spread diameter was 33 cm and 22 cm using the Abrams cone (see Figure 5).

Additionally, more tests were done using 450 Kg of cement with a water/cement ratio of 0.45 and 100 Kg of dredged material (mix C in Table 2). The percentage of arid was 56% and 44% of gravel. This third mix contains 1.7% of different admixture so the concrete obtained was extremely cohesive and self-levelling with a spread diameter of 56 cm. So, a self compacting concrete (SCC) was reached. Some pictures of this mix can be seen in Figure 6.

Step 6: Characterization of the SCC produced

4.3. Fresh state

The characterization of the fresh state of the SCC produced was performed by the determination of the consistency, air content, density and fluency. Results are given in Table 5.

4.4. Hardened state

The SCC produced with dredged sediments were tested after 28 days of curing in a humid chamber from a mechanical, environmental and durability point of view.

4.4.1. Mechanical tests

Compressive strength was measured over three concrete samples at the age of 28 days. Values obtained are those expected for a SCC developed with standard siliceous filler and they are given in Table 6.

4.4.2. Environmental tests

Even though the sediment was previously analysed and classified as Category I (31), leaching of the concrete samples, as concrete is the final material fabricated with the sediment, were carried out according EN 12457-2:2002 (40). This test is important in order to assure no leaching of hazardous metals in the final product.
Figure 5. Aspect of mix 1 using the sediment as filler.

Figure 6. Aspect of mix 2 using the sediment as filler: SCC.
The results obtained, concerning heavy metals, pH and conductivity in the leachate, are given in Table 7, where it can be seen that all data except nickel are below the detection limit of the analytical technique.

All the results of metal leaching expressed in Table 7 are under the limits established in the RD 1481/2001, by which regulates waste disposal by landfill, considering its last modification published in the order AAA/661/2013. The following table shows the lixiviation limit values for waste acceptable at landfills for inert waste (Table 8).

These values shall apply to waste acceptable at landfills for inert waste, calculated in terms of total liberation for the ratios between liquid and solid (L/S) of 10 l/kg and directly expressed in mg/l. Calculated using the test method UNE-EN 12457/Part 4 (L/S=10 l/kg, particle size<10 mm).

4.4.3. Durability tests

Parameters obtained by mercury intrusion porosimetry (total porosity, average pore diameter and bulk density) are given in Table 9, where it can be noticed that the concrete presents an averaged total porosity of 9.19 (% vol.), with a small average pore diameter (4V/A) and typical value of density for a SCC concrete.

Accumulated and differential pore size distribution of sample 1 is given in Figure 7, where it can be seen that the material presents a bi-modal distribution, with their highest maximum at around 0.08 µm and the other around 3 µm.

X-ray diffraction tests were carried out on grounded samples of concrete. The diffraction patterns obtained indicated that with the exception of the positive identification of a silicate of Fe and Mg, and a sulphate of Fe and Ti, no differences with a conventional concrete designed with normal silica filler are found with no evidence of any harmful phase for this material.

Other indicator of concrete durability is the electrical resistivity, which was measured using a commercial resistivimeter. Values obtained in two
specimens of concrete were, respectively, 49.4 KΩ cm and 49.15 KΩ cm, corresponding to a high durability concrete.

The transport properties through the matrix were measured by means of the determination of the steady and non-steady state chloride diffusion coefficients and the capillary absorption. The determination of the steady and non-steady state chloride diffusion coefficients is based on the accelerated measurement of the amount of chlorides arriving to the anolyte because of a voltage drop applied to the system.

Figure 7. Accumulated and differential pore size distribution of one of the samples of SCC (Mix C).

Figure 8. a) Schematic representation of the set up of the test. b) evolution of the accumulated Cl⁻ in the anolyte.
In Figure 8-a, schematic representation of the set up of the test is presented. In Figure 8-b the evolution of the accumulated Cl\(^-\) in the anolyte passing through the concrete is given.

From the representation in Figure 8-b, the effective and apparent diffusion coefficients can be determined. And from them, the binding factor of the material (alpha), supposed a linear binding. These results are given in Table 10.

The water take-up of hardened concrete was determined by capillary absorption. The water absorption of the two specimens of concrete, in percentage of weight, is presented as a function of the square root of the time in Figure 9, where it can be seen that the final absorption of the specimens is around 1.7% in weight of sample, which is quite low.

The summary of the parameters describing the capillary process is given in Table 11.

Values obtained for the resistant of water penetration, capillary suction and effective porosity are similar for both samples, and are indicative of materials with good resistance to the penetration of water.

Results obtained in the durability tests have been evaluated according to the durability indicators described in “Concrete design for a given structure service life” (45). They are presented in Table 12.

The low porosity of the concrete and its high electrical resistivity are indicators of a high durability.

However, values obtained in the measurement of the steady and non-steady diffusion coefficients indicate a moderated durability. Therefore, concerning environmental compatibility and durability, the main characteristics of the self compacting concrete developed are completely in agreement with those expected for a conventional concrete designed with normal silica filler.

5. CONCLUSIONS

In this paper, a feasibility protocol to assess the possibility of including a waste material in the production of a structural cementitious material has been designed. Then the protocol has been illustrated with its application to a real case: a dredged sediment of an Spanish harbour that has resulted feasible to be part of a self compacting concrete (SCC).

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REFERENCES


40. EN 12457-2:2002 Characterization of waste. Leaching. Compliance test for leaching of granular waste materials and sludges. One stage batch test at a liquid to solid ratio of 10 l/kg for materials with particle size below 4 mm (without or with size reduction).

41. UNE ENV 12506 Characterization of waste. Analysis of eluates. Determination of pH, As, Cd, Cr (VI), Cu, Ni, Pb, Zn, Cl\(^-\), NO\(_2\)\(^-\), SO\(_4\)\(^{2-}\).

42. ENV 13370:2001. Characterization of waste. Determination of Ammonium-N, AOX, conductivity, Hg, phenol index, TOC, CN\(^-\) easily liberatable and F\(^-\).


45. Spanish Code on Structural Concrete EHE-08.